

Preparation and characterisation of tamarind 4-hydroxybenzoic acid (THBA) resin and its use in extraction of heavy metal ions from industrial wastewater

AV Singh* and Indraj Kumar Kumawat

Department of Chemistry, Jai Narain Vyas University, Jodhpur – 342033, Rajasthan, India

Abstract

The adsorption of heavy metal cations, Pb(II), Cd(II), Cu(II), Zn(II) and Fe(II) from aqueous solution by newly-synthesized tamarind 4-hydroxybenzoic acid (THBA) resin was investigated. The resin was characterised on the basis of FTIR, elemental analysis, ion-exchange capacity and physico-chemical properties. The distribution coefficients (K_d) and percentage adsorption of metal ions on resin were determined by batch methods using atomic absorption spectrophotometry (AAS). The effect of experimental parameters, such as pH, treatment time, temperature, adsorbent dose, initial metal ion concentration and flow rate, on the removal of metal ions was also studied. THBA resin proved to be an effective adsorbent for the removal of different heavy metal ions from aqueous solution; removal efficiency followed the order: Fe(II) > Cu(II) > Zn(II) > Cd(II) > Pb(II). These results suggest that the cation exchange resin THBA holds great potential to remove cationic heavy metal species from industrial wastewater.

Keywords: batch method, distribution coefficient, ion-exchange capacity, FTIR spectra, THBA resin

Introduction

Toxic metals can be distinguished from other pollutants, since they are not biodegradable and can be accumulated in nature. They also cause various diseases and disorders when exceeding specific limits. An excess amount of heavy metal ions in aquatic environments is known to cause severe damage to human health and aquatic life. Therefore, separation of these metal ions from natural water at trace level is of paramount importance both for water purification and analysis. The complete removal of toxic heavy metal ions that are incompatible with biological systems requires expensive treatment in order to produce water that is again useful for domestic use (Monteagudo and Ortiz, 2000). It is essential to extend methods for removal of metal ions to decrease the pollution load on the environment. Classical techniques of heavy metal removal from solutions include the following processes: solvent extraction (Shukla and Rao, 2002), precipitation and coprecipitation (Kagaya et al., 2000), electrochemical reduction (Basha et al., 2008), chemical- and biosorption (Chang et al., 2002), pre-concentration (Ekic et al., 2004), reverse osmosis (Gibert et al., 2010) and ion exchange (Nabi and Naushad, 2008; Wong et al., 2007; Inamuddin et al., 2007; Khan and Alam, 2004).

In recent years, the adsorption process has also received much attention and has become one of the more popular methods for the removal of heavy metal ions from wastewater, because of its competitive and effective process. Numerous adsorbents have been reported for the removal of toxic metal ions, such as chitin (Santosa et al., 2007), chitosan (Wan Nagh et al., 2002), starch (Delval et al., 2005), cellulose

(El-Mensawy et al., 2008; Gupta et al., 2002; Gerente et al., 2000; Zhou et al., 2004; Guclu et al., 2003), guaran (Singh et al., 1996), and cyclodextrin (Kozłowski et al., 2005), which are not only eco-friendly and cost-effective but are also effective in remediation of common effluents present in wastewater. Other polysaccharide-based materials and aluminosilicates are used as adsorbents in wastewater treatment (Crini, 2005; Aguiar et al., 2002; Rubio et al., 1997). Adsorption using commercial activated carbon (CAC) can remove metal ions from wastewater, such as Cd(II), Ni(II), Cr(III) and Cu(II) (Ramos et al., 1997; Shim et al., 2001; Ouki et al., 1999; Monser et al., 2002). However, CAC is an expensive material. Resins prepared with divinyl benzene-styrene backbone are hydrophobic (Sarin et al., 1980). These resins, prepared from petrochemicals, are costly. Their ion exchange capacity depends mainly on the quantity of functional groups and the pH of the solution. The most widespread chelating functional groups used for removal of metal ions from effluents are thiol (Saha et al., 2000), thiourea (Hubicki et al., 2009), dithionite (Grote et al., 2001), aspartate (Chuh-Yean et al., 2008) and triisobutyl phosphine sulphide (Hidalgo et al., 2001).

Tamarind kernel powder (TKP) is obtained from the seeds of the tamarind tree, *Tamarindus indica*, a common forest and cultivated tree found primarily in India, Burma, Bangladesh and Sri Lanka. TKP is a polymer with an average molecular weight of 52 350 and a monomer containing 3 sugars, glucose, galactose, and xylose in a molar ratio of 3:1:2, was reported by Khanna and Sarin (1987). The polymer consists of a cellulose-type spine, which carries xylose and galactoxylose substituents. About 80% of the glucose residues are substituted by a 1+6 linked xylose units, which themselves are partially substituted by 1-2 galactose residues (Lang et al., 1993). The structure of tamarind kernel powder is given in Fig. 1.

The ion exchanger based on tamarind kernel powder is hydrophilic and biodegradable, whereas ion exchangers prepared from petrochemical products are hydrophobic and not

* To whom all correspondence should be addressed.

☎ +91 9461286995; fax: +91 0291 2720436;

e-mail: reshvikram04@rediffmail.com

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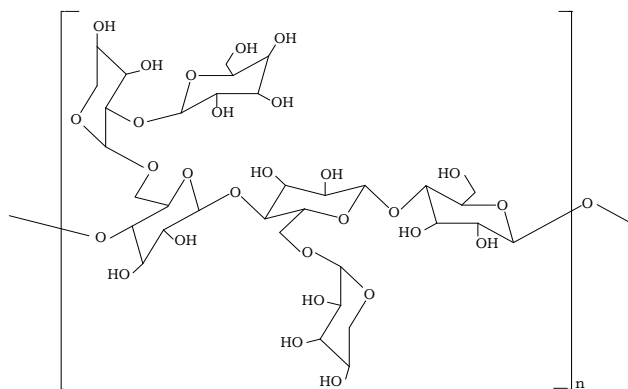


Figure 1
Structure of tamarind

biodegradable. Due to rising prices of petroleum products TKP has been selected for development of tamarind 4-hydroxybenzoic acid (THBA) resin. Its cost is low and it is locally available in large quantities from agricultural resources; such biopolymers are environmentally friendly.

The work reported herein was undertaken to study the synthesis and characterisation of THBA resin and its applications for removal and recovery of toxic metal ions from the effluent of Paradise Steel Industry, Jodhpur, India, in the form of batch and column processes. The influence of various factors affecting adsorption conditions, pH, temperature, stirring time, treatment time and resin dose, on the removal of metal ions was studied.

Materials and methods

Chemicals

Analytical grade chemicals were used in all experiments. All compounds used in the synthesis were of commercial high purity grade. Tamarind kernel powder (Ases Chemical Works, Jodhpur, India), epichlorohydrin (British Drug Houses Ltd. Poole England), 4-hydroxybenzoic acid (Loba Chemic Pvt Ltd, Mumbai, India), sodium hydroxide (Sarabhai M. Chemicals, Baroda, India), dioxane (E Merk, Mumbai, India), methanol (Allied Chemical Corporation, Varodra, India), hydrochloric acid (CDH Pvt. Ltd, New Delhi). The standard solution (self-made) was a 1 000 mg/kg solution of metal ions ($\text{CuSO}_4 \cdot 5\text{H}_2\text{O}$, $(\text{NH}_4)_2\text{Fe}(\text{SO}_4)_2 \cdot 6\text{H}_2\text{O}$, $\text{CdSO}_4 \cdot \text{H}_2\text{O}$, $\text{Pb}(\text{NO}_3)_2$, $\text{ZnSO}_4 \cdot 7\text{H}_2\text{O}$ (Loba Chemic Pvt Ltd, Mumbai, India), which was then diluted to different solution concentrations. The characteristic features of the effluent of Paradise Steel Industry, Jodhpur (Rajasthan) are presented in Table 1.

Synthesis of tamarind 4-hydroxybenzoic acid (THBA) resin

Step A: Preparation of epoxypropyl ether of tamarind

An amount of 182.2 g of tamarind kernel powder (0.2 mol) was taken in a round-bottom flask and was slurried with dioxane. Subsequently, 15 ml of 40% (w/v) sodium hydroxide was added to make the solution alkaline, until the pH reached 8.5. The solution was stirred at 60°C. Then 9.25 g (0.1 mol) epoxychloropropane (epichlorohydrin) was added with constant stirring. The stirring was further continued for 5 h, at 60°C. The product, epoxypropyl ether of tamarind, was filtered under vacuum, washed with methanol to remove impurities, and dried (Singh et al., 2011).

Step B: Preparation of tamarind 4-hydroxybenzoic acid (THBA) resin

Epoxypropyl ether of tamarind, synthesized in Step A, was allowed to react with 13.8 g (0.1 mol) of 4-hydroxybenzoic acid; the mixture was stirred for 5 h at 60°C and left overnight. The product was filtered under vacuum and washed with 90% methanol, which also contained a few drops of hydrochloric acid to remove inorganic impurities. Finally, it was washed with pure methanol. The product, tamarind 4-hydroxybenzoic acid (THBA) resin, was a free-flowing light yellow powder and the yield was 205.3 g. The reaction scheme is shown in Fig. 2.

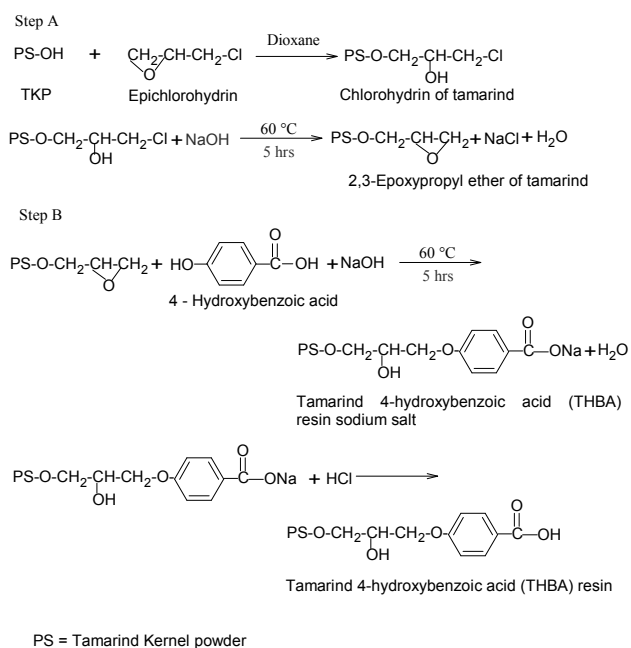


Figure 2
Synthesis of tamarind 4-hydroxybenzoic acid (THBA) resin

Appearance: Turbid							
Colour: Dirty brown							
pH: 6.2							
Total hardness: 850.9 mg/kg							
Metal ion	Fe ²⁺	Cu ²⁺	Zn ²⁺	Pb ²⁺	Cd ²⁺	Mg ²⁺	Ca ²⁺
Concentration (in ppm)	1.12	0.52	5.18	0.69	0.18	102.6	172.9
Other anions (mg/kg): fluoride = 0.72; sulphate = 468; cyanide = 0.04							
Acceptable level of toxic metal ions in water (mg/kg): Fe ²⁺ = 0.1, Cu ²⁺ = 0.05, Zn ²⁺ = 5.0, Pb ²⁺ = 0.1, Cd ²⁺ = 0.1							

Water washing of resin

The resins were washed with diluted HCl to create the hydrogen form, and the hydrogen form of the resin was then washed with double-distilled water to remove all of the excess acid. The dried material at 378 K was used for further experimental work.

Physico-chemical characteristics

The physico-chemical properties, such as percentage moisture content, percentage solid, true density, apparent density, void volume fraction and volume capacity and ion exchange capacity were analysed according to standard methods (Vogel, 1989; Helfferich, 1962). The physico-chemical characteristics of the synthesised tamarind 4-hydroxybenzoic acid are summarised in Table 2.

Property	Value (SD)
Active group	-COOH
Ionic form as shipped	H ⁺
% moisture	13.7 ± 0.5
% solid	86.3 ± 0.5
True density (g/cm ³)	1.31 ± 0.02
Apparent density (g/ml)	0.6632 ± 0.0052
Void volume fraction	0.3571 ± 0.1131
Volume capacity (mmol/cm ³)	2.84 ± 0.05
Total exchange capacity (meq/g)	4.62
Particle size (mm)	0.26 ± 0.05
Thermal stability	404 ^o
Operating pH range	1-7

Batch experiment

The distribution coefficients (K_d) and percentage adsorption of metal ions on resin were determined using the batch method. The pH of solution was adjusted by a suitable buffer at desired values: KCl-HCl buffer for pH 1 to 2; acetate buffer for pH 3 to 7. A sample solution (100 ml) containing a known concentration of the studied metal ion was transferred to an Erlenmeyer flask and, after adjusting the pH, 100 mg of the modified THBA resin was added to the solution, the mixture was shaken continuously in a temperature-controlled shaker at 25 ± 2°C for 4 h and the contents were equilibrated. The solution was filtered through Whatman filter paper no. 40. The residue on the filter paper was equilibrated with 0.1–2.0 mol/m³ HCl and the solution was filtered through Whatman filter paper no. 42.

The toxic metal ion concentration in the filtrate as well as in the residue was estimated using a Perkin-Elmer model 2380 atomic absorption spectrophotometer. The calibration curves for different metal ions were plotted, by analysing a series of standard solutions of metal ions using AAS. The flame type was air-acetylene and absorption wavelengths were Zn²⁺ (213.6 nm), Fe²⁺ (324.8 nm), Cu²⁺ (324.8 nm) Cd²⁺ (228.8 nm) and Pb²⁺ (283.2 nm). The concentration of metal ion in the filtrate was determined by the calibration curves. The distribution coefficient (K_d) and percentage adsorption of the metal ions on THBA resin were calculated using the equations:

$$K_d = \frac{\text{Amount of metal ion in resin phase}}{\text{Amount of metal ion in solution phase}} \times \frac{V}{m} \text{ ml/g} \quad (1)$$

where:

V is the volume of the solution (ml)

m is the weight of the dry resin (g)

$$\text{Percentage adsorption of metal ions} = \frac{C_i - C_f}{C_f} \times 100 \quad (2)$$

where:

C_i is the initial concentration of metal ion in solution and

C_f is the final amount of metal ion in solution after equilibrium with resin.

Column experiment

In the column experiment, a glass tube with 1.6 cm internal diameter and 20 cm height, packed with 8 cm of resin (7.0 g), was used. 50 ml of the sample metal ion solution was passed through the column at a flow rate of 2 ml/min. The flow rate was controlled by a peristaltic pump. The column was washed with 20 ml of deionised water and the washing was rejected. The metal ions were eluted quantitatively with different strengths of acids.

Ion exchange capacity (IEC) determination

Resin capacity is usually expressed in terms of equivalents per litre (eq/l) of resin or milli-equivalents per dry gram of resin. The ion exchange capacity, which is generally taken as a measure of the hydrogen ion liberated by a neutral salt flowing through the composite cation exchanger, was determined by a standard column process. 0.1 g (dry mass) of the composite ion exchange material in H⁺ form was placed in a glass column with a glass wool support at the bottom. It was washed with demineralised water to remove any excess acid that remained on the particles. The hydrogen ions were eluted with 0.1 M solution of different alkali and alkaline earth salts. The flow rate was kept at 4 ml/min. The collected effluent was titrated against a standard solution of sodium hydroxide using phenolphthalein as an indicator. The hydrogen ions released were then calculated according to the method described by Vogel (1989) and was found to be 4.62 meq/g.

Results and discussion

Fourier transform infrared spectroscopy (FTIR) characterisation of tamarind 4-hydroxybenzoic acid (THBA) resin

A Varian model 640 FTIR instrument was employed for the FTIR spectral analysis of synthesised THBA resin using KBr pellets. The FTIR spectrum of tamarind powder shows a broad band in the region 3 600 – 3 200 cm⁻¹ characteristic of -OH stretching. The peak at 2 936 cm⁻¹ is attributed to C-H stretching vibrations; a strong and sharp peak at 1 650 cm⁻¹ is due to O-H bending and the variable peak at 1 480 – 1 350 cm⁻¹ is attributed to C-H bending. A strong peak at 1 300 – 1 000 cm⁻¹ denotes C-O stretching vibration. The FTIR spectrum of tamarind is shown in Fig. 3.

The FTIR spectrum of THBA resin shows a peak at 3 100 – 3 000 cm⁻¹ which is attributed to C-H stretching vibration in aromatic nuclei. A strong peak in the region 1 250 – 1 000 cm⁻¹ denotes C-O stretching vibrations. The peak at 1 700 – 1 680

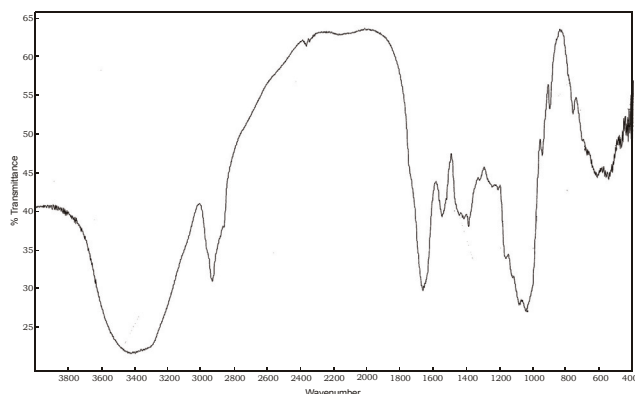


Figure 3
FTIR spectrum of tamarind kernel powder

cm^{-1} is attributed to C=O stretching vibrations of aryl carboxylic acid. The peaks at 1620–1450 cm^{-1} are attributed to C=C stretching in aromatic nuclei. The THBA resin in H^+ form shows symmetric stretching vibration in the region 2500–3000 cm^{-1} , which is attributed to the –OH group in carboxylic acid. The spectral peaks for polysaccharides are generally observed in the region 3600–3200 cm^{-1} due to O–H stretching frequency (Rao, 1963). The FTIR spectrum of THBA resin is given in Fig. 4.

Elemental analysis

A Carlo Erba model 1160 elemental analyser was used for determination of the mass percentage of elements in the synthesized resin. This was theoretically calculated as (%) 47.73 (C); 5.97 (H); and 46.29 (O), and analysed to be (%) 47.36 (C); 5.14 (H); and 45.78 (O). The results of the elemental analysis are in good agreement with the calculated value.

Distribution coefficient (K_d) of metal ions in effluent of Paradise Steel Industry, Jodhpur

The pH has a strong effect on the distribution coefficient (K_d) of the metal ions. Table 3 shows the variation of K_d values of different metal ions with (H^+), which reveals the increase of the K_d values with decreasing acidity of the aqueous solution and optimum results obtained at pH range 4–6. Metal sorption starts when the pH rises to the range where most acidic ion exchange sites start to exchange hydronium ions for metal ions, and the capacity reaches the maximum value in the pH range where all the ion exchange sites take part in the reaction and the functional group is able to form complexes with the metal cations (Nenov and Bonev, 2010). The decrease in K_d values after the maximum in the neutral and alkaline region can be explained by the complex formation of THBA resin with heavy metal ions. High values of the distribution coefficient indicate that the metal has been retained by the resin through adsorption, while low values of K_d indicate that a large fraction of the metal remain in solution.

Removal of metal ions from effluent of Paradise Steel Industry, Jodhpur, by THBA Resin

The results for percentage removal of metal ions from effluent of Paradise Steel Industry, Jodhpur, by THBA resin are summarised in Table 4. The adsorption percentage for the

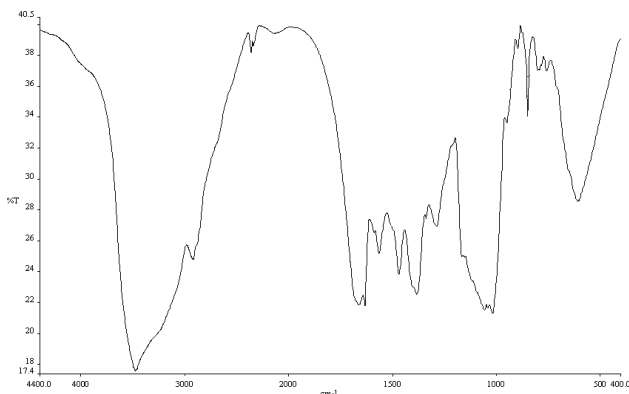


Figure 4
FTIR spectrum of tamarind 4-hydroxybenzoic acid (THBA) resin

Table 3
Distribution coefficient (K_d) of metal ions from effluent of Paradise Steel Industry, Jodhpur, on THBA resin ($K_d \times 10^2$)

pH	Fe ²⁺	Zn ²⁺	Cu ²⁺	Cd ²⁺	Pb ²⁺
1	19.63	11.06	5.34	6.20	12.55
2	31.66	16.85	7.68	11.02	19.35
3	72.65	73.81	16.75	15.16	32.32
4	219.91	130.87	40.09	34.39	54.73
5	56.86	43.57	83.71	51.18	100.89
6	15.71	15.92	164.58	73.78	45.05
7	8.43	13.04	14.86	17.22	17.80

metal ions on the THBA resin increases with pH and reaches its maximum value at pH 4–6, declining at higher pH. It has been found that the maximum removal percentages for Fe(II), Cu(II), Zn(II), Pb(II) and Cd(II) are 95.65%, 94.27%, 92.90%, 90.98% and 87.74%, respectively.

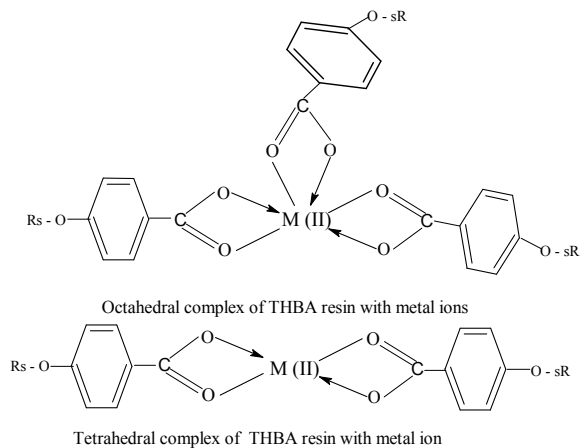
It is clear from Table 4 that the percentage removal of metal ions first increases and then decreases with increasing pH. Similar results have been reported for a previous study with amberlite IR-120 (Demirbas et al., 2004). This suggests that selectivity of the metal ion is dependent on pH. The maximum removal of toxic metal ions is obtained when a strong electric field is present and the electrostatic effect may become the dominant factor, such that small ions, which have a huge charge density, are bound more strongly with the resin (Gode and Pehlivan, 2005). In an acidic medium, the H^+ ions of the THBA resin easily exchange with metal ions.

Table 4
Percentage removal of metal ions from effluent of Paradise Steel Industry, Jodhpur, by THBA resin

pH	Fe ²⁺	Zn ²⁺	Cu ²⁺	Cd ²⁺	Pb ²⁺
1	66.25	52.73	34.71	38.27	55.66
2	76.00	62.76	43.44	52.41	65.93
3	87.90	88.06	62.62	60.26	76.37
4	95.65	92.90	80.03	77.47	84.55
5	85.08	81.33	89.32	83.65	90.98
6	61.11	61.41	94.27	87.74	81.33
7	45.71	56.60	59.78	63.92	64.40

Mechanisms for adsorption of metal ions on THBA resin

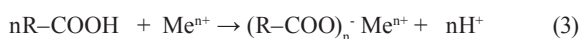
The heavy metal ions' adsorption onto the surface of the THBA resin is assumed to occur through ion-exchange and



Rs = Resin matrix

Figure 5
Complex structure of THBA resin

chelation between charged heavy metal ions and ionised carboxylic groups, which is depicted in the following equation:



Hydrogen ions in the carboxylic group ($-COOH$) of the resin can serve as exchangeable ions with metal cations. The THBA resin has many carboxyl ions ($-COOH$), which can serve as the adsorption sites for ion exchange and chelation. In an aqueous solution these groups within the resin dissociate to yield protons, leading to a decrease in solution pH. This is testified to by the pH variation of the suspension before and after adsorption, as stated above. This suggests that, along with the shift or disappearance of some characteristic group, chelation occurs between the adsorbent and adsorbate. In consideration of this, it can be deduced that ion exchange and chelation coexists during the entire adsorption process. The complex structure of THBA resin is given in Fig. 5.

Statistical analysis

The relative standard deviation values (RSD) of optimum removal percentages for the metal ions are shown in Table 5. All data represent the mean of the results of 3 independent experiments. The results revealed that the relative standard deviation (RSD) of the method was lower than 3.0, which indicated that the method had good precision for the analysis of trace metal ions in solution samples.

Metal ions	Amount of metal ions in effluent (mg)	Amount loaded on THBA resin (mg)	Removal percentage	RSD %
Fe ²⁺	1.12	1.071	95.65	1.62
Cu ²⁺	0.52	0.492	94.27	1.32
Zn ²⁺	5.18	4.812	94.27	1.46
Pb ²⁺	0.692	0.629	90.98	2.64
Cd ²⁺	0.186	0.163	87.74	2.16

Use of eluent and recovery of metal ions

The recovery of metal ions was determined in the column experiment. In the column experiment the metal ions were eluted quantitatively with different strengths of acid. The Cd(II) was eluted with 0.1 mol/m³ HCl, Pb(II) with 0.5 mol/m³ HCl, Zn(II) with 1.0 mol/m³ HCl, Cu(II) with 1.5 mol/m³ HCl and Fe(II) was eluted with 2.0 mol/m³ HCl. The resin column was then washed thoroughly with demineralised water. The amounts of metal ions in the filtrate solution were analysed using AAS. Recoveries of Cd(II), Pb(II), Zn(II), Cu(II) and Fe(II) were determined as 96.31%, 96.66%, 97.36%, 98.17% and 99.06%, respectively. The elution of metal ions was carried out with hydrochloric acid solution, using the advantage that the chloride ion is an acceptable matrix for both AAS and spectrophotometric determination of metal ions. Data presented in Table 6 indicate that different quantities and strengths of hydrochloric acid solution could afford quantitative elution of different metal ions from the resin. Dev and Rao (1996) used modified Amberlite XAD-4 with bis-(N,N-Salicylidene)-1,3-propanediamine for the separation of Ni(II), Fe(II), Ni(II), Co(II), Zn(II), Hg(II), Pb(II), and sorbed metals were eluted by 1 M HCl with about 100% recovery. The recoveries of metal ions by elution were calculated according to Eq. (4).

$$\text{Recovery \%} = \frac{\text{Amount of metal ions desorbed to the elution medium}}{\text{Amount of metal cations adsorbed on the resin}} \times 100 \quad (4)$$

Metal ion	Amount loaded (mg)	Amount found (mg)	Recovery percentage	Eluent HCl (mol/m ³)	Eluent volume (ml)
Fe ²⁺	1.071	1.061	99.06	2.0	60
Cu ²⁺	0.492	0.483	98.17	1.5	55
Zn ²⁺	4.812	4.685	97.36	1.0	50
Pb ²⁺	0.629	0.608	96.66	0.5	45
Cd ²⁺	0.163	0.157	96.31	0.1	40

Column reusability

It was observed that the adsorbency of different metal ions on the THBA resin after 10 cycles (adsorption and desorption), and the ion exchange capacities of the resin were almost constant. The adsorbed metal ions were easily desorbed by treatment with different strengths of acid, at room temperature. Table 7 shows the adsorbency of different metal ions on the THBA resin after 10 cycles (adsorption and desorption).

Metal ions	Adsorption percentage for different metal ions on THBA resin after a number of cycles				
	1 cycle	2 cycles	4 cycles	8 cycles	10 cycles
Fe ²⁺	95.65	93.54	93.46	93.45	93.45
Cu ²⁺	94.27	92.21	92.14	92.13	92.13
Zn ²⁺	92.90	90.68	90.42	90.41	90.40
Pb ²⁺	90.98	88.78	88.64	88.63	88.62
Cd ²⁺	87.74	85.58	85.44	85.42	96.42

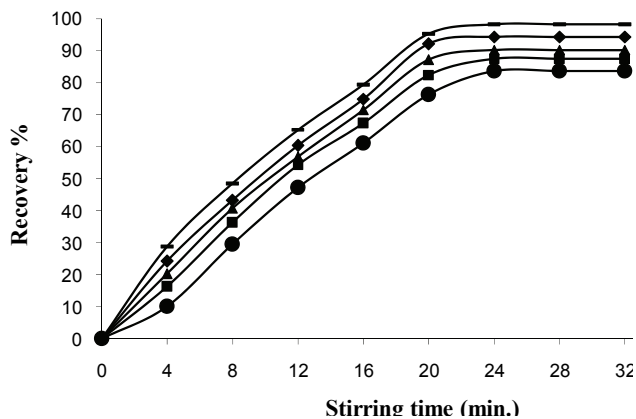


Figure 6
Effect of stirring time on the recovery of metal ions

Effect of stirring time on the recovery of metal ions

To determine the rate of sorption of metal ions on the THBA resin, batch experiments were elaborated by shaking the solution containing the ion with 100 mg of the THBA resin at room temperature (25°C). Aliquots of 1.0 ml solution were taken for analysis at pre-determined intervals. The concentration of metal ion in the supernatant solution was determined and the amount of metal ion sorbed on the THBA resin was calculated by mass balance. The sorption half-time ($t_{1/2}$), defined as the time needed to reach 50% of the total sorption capacity, was estimated from Fig. 6. From the data obtained it was observed that the maximum sorption of metal ions with THBA resin reached its equilibrium time after about 24 min. However, the time required for 50% sorption of metal ions was 11 min for THBA resin.

Effect of pH

The pH is an important parameter for adsorption of metal ions from aqueous solution because it affects the solubility of the metal ions, concentration of the counter ions on the functional groups of the adsorbent and the degree of ionisation of the adsorbent during reaction. To examine the variation of adsorption percentage for metal ions with pH, the pH was varied from 1.0 to 7.0 as given in Table 4. The uptake of free metal ions depends on pH, where optimum adsorption of metal ions occurs at different pHs (ranging from 4 to 6), and then declines at higher pH. Adsorption of metal ions on THBA resin increased over a pH range from 1.0 to 6.0, then decreased. At lower pH (acidic pH), the adsorbent surface will be completely covered with hydronium ions which compete strongly with metal ions for adsorption sites. With an increase in pH, the concentration of H_3O^+ ions decreases, facilitating the adsorption of metal ions by the adsorbent (Radovic et al., 1997).

Effect of contact time

Treatment time indicates that the adsorption percentage for metal ions increased with an increase in contact time before equilibrium is reached. It is clear from Fig. 7 that adsorption of metal ions on THBA resin increased when contact time was increased from 30 to 210 min; optimum contact time for THBA adsorbent was found to be 210 min. Other parameters, such as pH of solution and agitation speed, were kept optimum,

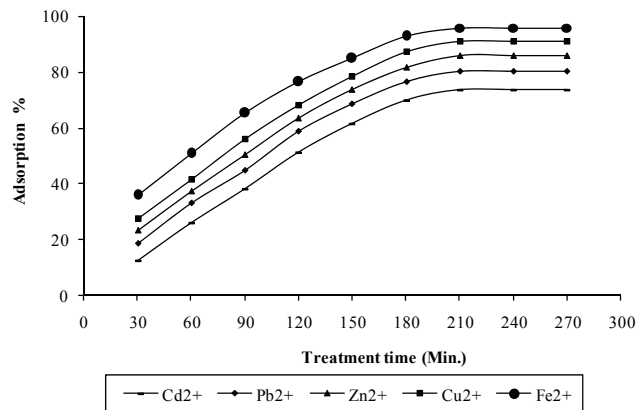


Figure 7
Effect of changing treatment time on the percentage adsorption of different metal ions on THBA resin

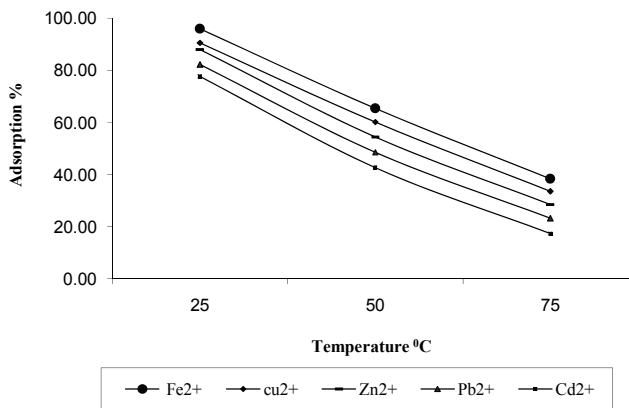


Figure 8
Effect of changing treatment temperature on the percentage adsorption of different metal ions on THBA resin

while temperature was kept at 25°C. Greater availability of carboxylic and ether functional groups on the surface of TKP, which are required for interaction with metal ions, significantly improved the binding capacity and the process proceeded rapidly. This result is important, as equilibrium time is one of the important parameters for an economical wastewater treatment system.

Effect of treatment temperature

The effect of temperature on the adsorption percentage for metal ions onto THBA resin was studied. Figure 8 shows the effect of treatment temperature. The percentage adsorption of metal ions decreases by increasing the treatment temperature from 25°C to 50°C and then 75°C, at the optimum treatment time of 4 h. The working capacity of an ion exchanger depends on metal concentrations and temperatures. This observation is in full agreement with results published by Khalil et al. (1998).

Effect of THBA dose on adsorption of metal ions

The resin amount is also an important parameter influencing the quantitative adsorption of metal ions. The retention of the metal ions was examined in relation to the amount of resin used, in the range of 20–160 mg and equilibrated for 4 h. It is apparent that by increasing the resin amount, the adsorption

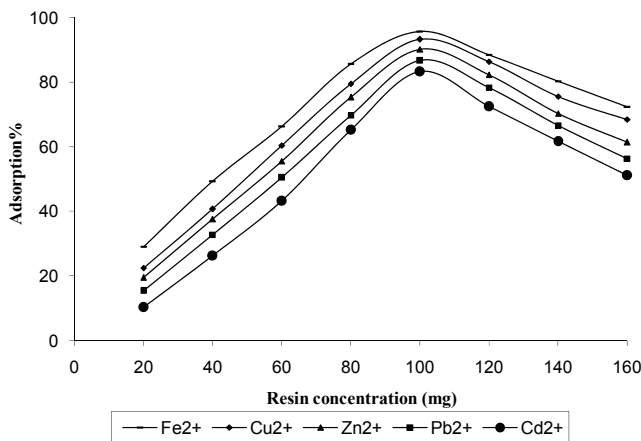


Figure 9
Effect of adsorbent dose on adsorption of metal ions on THBA resin

density and the amount of metal ions adsorbed per unit mass increases. The maximum adsorption by THBA resin was achieved with an adsorbent dose of 100 mg and continued decreasing up to 160 mg. The initial increase in adsorption percentage for the metal ions was due to the greater surface area and increased number of adsorption sites available (Rengaraj et al., 2008). On increasing the THBA resin concentration further, the binding of metal ions steadily decreased. This effect might be attributed to overlapping or aggregation of adsorption sites of resin resulting in a decrease in the total surface area of the adsorbent. The results are shown in Fig. 9.

Conclusions

The removal of Fe(II), Cu(II), Zn(II), Pb(II) and Cd(II) ions by THBA resin is considered to be a very promising technique, due to its cost effectiveness, environmental friendliness, and short duration. The optimum conditions for removal of Fe(II), Cu(II), Zn(II), Cd(II) and Pb(II) ions using THBA cation exchange resin were found to be pH 4–6 and stirring time ~25 min. A concentration of 0.1–2.00 mol/m³ HCl as eluent was sufficient to obtain maximum recovery for all metal ions. The relative standard deviations (RSD) for all of the experiments, indicating their repeatability and reproducibility were less than 3.0%. These results indicate that the present method can be used for quantitative analyses and removal of toxic metal ions. The exchange bed could be used for more than 10 cycles with little loss of exchange capacity. The selectivity and ion-exchange capacity of these materials towards metals ions can be controlled by the pH of the medium, stirring time, contact time, temperature, adsorbent dose, etc. Therefore the THBA resin is applicable for the removal and recovery of heavy metal ions from industrial effluents.

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